Original Research

Characterization of Nitro-Aromatic Compounds of Ethylbenzene Secondary Organic Aerosol with Acidic Sodium Nitrate Fine Particles

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Abstract

Acidic sodium nitrate (NaNO₃) is the fine particles commonly found in urban atmosphere, which is beneficial to the formation of nitro-aromatic compounds (NACs). The NACs and optics of secondary organic aerosol (SOA) generated from the photo-oxidation of ethylbenzene with acid NaNO₃ fine particles were characterized in the current study. The results demonstrated that acidic NaNO₃ can significantly promote the formation of SOA. Compared with the carbonyls detected in SOA without fine particles, nitrophenols, nitrocatechols and their derivatives are measured as the major components of SOA with acid NaNO₃. The N=O chromosphore in NACs enhanced the light absorption capacity of SOA. The mass absorption coefficient averaged over 200-600 nm (<MAC>) of SOA gradually increased with the increasing of acidic NaNO₃. While, the <MAC> of SOA with 100 μ g/m³ acidic NaNO₃ was comparable to that of biomass burning organic aerosols. These would provide the experimental basis for studying the formation and optics of anthropogenic NACs in the background of high concentration of acidi NaNO, fine particles.

Keywords: nitro-aromatic compounds, ethylbenzene, secondary organic aerosol, sodium nitrate, fine particles

Introduction

Nitro-aromatic compounds (NACs), especially for the -OH and $-NO_2$ groups attached to the aromatic

ring, such as nitrophenols, nitrocatechols and their derivatives, are an important class of organic nitrogen of aerosol particles in atmosphere [1-2]. NACs have been detected in urban and rural atmosphere, with the concentrations ranged from 0.1 to 147.4 ng/m³ [2-4]. The observational field experiments demonstrated that biomass combustion contributed about 10-21% of the NACs in aerosol particles [5-6]. Also, NACs can

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be secondarily formed via OH-initiated oxidation of aromatic compounds released by anthropogenic emission sources in the presence of NO_x [7-8]. As suggested by Ji et al. [9], aromatics reacted with OH radicals result in the formation of hydroxycyclohexadienyl radical via the major OH addition, which subsequent reaction with O₂ lead to the formation of phenolic compounds under atmospheric condition. Phenolic products can further initiated by OH radical to form hydroxyphenoxy radicals, which then react with NO₂ to generate NACs [7-10]. Besides to their mutagenesis and genotoxicity to humans, NACs with strong chromophore of N=O were responsible for the light absorption by brown carbon (BrC), contributing to radiative balance and visibility degradation [11-12]. Thus, the study of the chemical composition and formation process of NACs has great atmospheric significance.

Ethylbenzene and other aromatic compounds are expected to be vital precursors of NACs in urban atmospheres [13-15]. OH-initiated photo-oxidation of aromatic compounds in the presence of NO_x leading to the formation of nitrophenols, dinitrophenols and other NACs, which result in the formation of secondary organic aerosol (SOA) particles [7-10, 13-15]. Compared with the gaseous formation of NACs, aqueous aromatic nitration is another formation pathway of NACs [16-17]. For example, nitrophenols can be formed via the hydroxylation and nitration of benzene upon UV irradiation of nitrite in aqueous solutions [16]. Some of the aromatic compounds emitted to the atmosphere remain in the gas phase were oxidized by OH radicals and NO₂ to produce NACs. The other part entered into the aqueous phase such as liquid aerosol particle, water droplet, and cloud in the atmosphere. When nitrite exists in the aqueous phase, NACs can be generated under ultraviolet light [8, 18]. However, the aqueous formation pathway of NACs is less well-understood.

Owing to the development of China's economy, the number of motor vehicles is increasing, the use of fossil fuels in large quantities, and the urban atmosphere contains relatively large concentrations of acidic inorganic fine particles [19-20]. For example, the concentration of Na⁺, H⁺, NH₄⁺, NO₃⁻ and other inorganic ions in fine particles achieved 100-300 µg/m³ during hazy days in Beijing and Shanghai [21-23]. Therefore, there is high concentration of acid sodium nitrate (NaNO₃) fine particle in Chinese urban atmosphere [19, 20]. These fine particles have large specific surface area, which can become the center of condensation and heterogeneous reactions of semi-volatile compounds [24]. In addition, the surface of fine acidic NaNO, particles absorbs moisture to form the liquid aerosol, and the H⁺ and NO₂⁻ can generate OH and NO₂ radicals via photolysis upon UV irradiation [25], which is beneficial to the formation of NACs. Therefore, the study of the influence of high concentration of acidic NaNO₂ fine particles on the formation of aromatic SOA and the composition of NACs is helpful to understand the aqueous formation mechanism of NACs. Recently, our group has performed the influence of inorganic fine particles such as ammonium sulfate and calcium chloride on the chemical composition of toluene, 1,3,5-trimethylbenzene in smog chamber [26-28]. So, PM_{2.5} particle detector, UV-Vis spectrophotometer (UV-Vis) and liquid chromatography-mass spectrometry (LC-MS) were utilized to measure the ethylbenzene SOA of different concentrations of acidic NaNO₃ fine particles in this study. The influence of acidic NaNO₃ fine particles on the NACs composition and optical properties of ethylbenzene SOA were discussed. These would provide experimental basis for studying the formation mechanism of NACs under the background of acidic inorganic fine particles.

Experimental

Materials

Ethylbenzene (>99%) and hydrogen peroxide (30%) were obtained from Sigma-Aldrich Chemistry Corporation, Germany. Methanol (>99%), nitric acid (65-68%) and sodium nitrate (99.9%) were supplied by the Third Reagent Factory of Tianjin.

Acidic inorganic fine particles were produced by aspirating the 0.001 mol/L NaNO₃ and HNO₃ mixed solution via the atomizer (TSI Inc. Model 3076, USA), passed through a dryer (TSI Inc. Model 3062, USA) and a neutralizer (TSI Inc., Model 3054, USA), and then introduced into the chamber [26-28]. The diameter of the acidic NaNO₃ fine particles was less than 2 μ m, while the concentration was established at a certain value through controlling the filling time of the fine particles into the chamber.

Ethylbenzene SOA Formation without and with Acidic NaNO₃ Fine Particles

Formation of ethylbenzene SOA particles without and with different concentration of acidic NaNO, fine particles was performed in home-made chamber [26-28]. Ethylbenzene, hydrogen peroxide and acidic NaNO₂ fine particles were delivered to chamber, which was then filled with zero air to full volume. The concentration of ethylbenzene and hydrogen peroxide was fixed as 1 ppm and 10 ppm, respectively. 8 experiments were carried out in the absence and presence of acidic NaNO, fine particles in the concentration of 25, 50, 75, 100, 150, 200 and 300 µg/m³, respectively. The temperature and relative humidity in chamber was about 298 K and 37%. Four UV lamps were used to irradiate hydrogen peroxide to form OH radicals [29], which initiated the oxidation of ethylbenzene to generate SOA. The concentration of ethylbenzene and SOA particles in the chamber was detected using GC-FID (Agilent 7820A, USA) and PM₂₅ particle detector (Shenzhen Huashengchang Machinery Industry Co., Ltd, CEM-DT96, China). Each experiment was performed for 3 times, the averaged results were utilized as the final data.

Chemical Characterization of NACs in Ethylbenzene SOA Using UV-Vis and LC-MS

After photo-oxidation, the ethylbenzene SOA was collected onto the polytetrafluoroethylene membrane filter and extracted into 2% methanol solution using sonication. The UV-Vis spectra of the extract solutions were measured by double beam UV-6100S spectrophotometer (Mapada Instruments, China). Also, extract solutions were detected by LC-MS equipped with electrospray ionization (ESI) (Agilent-1200, Agilent-6320, USA). The mobile phase was methanol and ultra-pure water (1:1 v/v) with the rate of 0.20 μ L/min. Each sample was detected without LC column, and measured in the range of 50-1000 amu of the negative mode [30].

Optical Characterization of Ethylbenzene SOA

Mass absorption coefficient (MAC) is often utilized to characterize the optical properties of aerosol particles. Aerosol samples were collected by membrane sampler, and the soluble components in the particles were extracted by water or methanol. The absorbance was obtained by measuring the light attenuation after the light passed through the extract solution with the help of ultraviolet visible spectrophotometer. The MAC parameters were obtained according to the length of the absorption cell and the concentration of organic matter in the sample [31-32]:

$$MAC(\lambda) = \frac{A^{\text{solution}}(\lambda) \times \ln 10}{b \times C_{\text{mass}}}$$
(1)

$$\langle MAC \rangle = \frac{1}{(\lambda_2 - \lambda_1)} \times \int_{\lambda_1}^{\lambda_2} MAC(\lambda) d\lambda$$
 (2)

...where A^{solution} in formula (1) is the absorbance of the extract at a specific wavelength (λ), ln10 is a constant, b is the optical path length, and C_{mass} is the mass concentration of SOA organic carbon in the extract solution. According to formula (2), the averaged MAC value <MAC> is obtained by integrating in the measurement wavelength range (λ_1 - λ_2). The absorption spectrum and total organic carbon concentration of the ethylbenzene SOA extract solutions at 200-600 nm were measured by UV-6100S spectrophotometer and total organic carbon analyzer (TOC-L, Shimadzu Company, Japan). The mass absorption coefficient averaged over 200-600 nm (<MAC>) of ethylbenzene SOA was obtained according to formula (1) and (2).

Results

Formation of Ethylbenzene SOA without and with Acidic NaNO, Fine Particles

The formation of ethylbenzene SOA without fine particles was measured firstly. For the mixed gases of 1 ppm ethylbenzene and 10 ppm hydrogen peroxide, the concentration of SOA with irradiation time measured by PM25 particle detector was shown in Fig. 1. In the first 15 min, few particles (4 μ g/m³) were detected by PM2,5 particle detector. OH radicals formed from the photolysis of hydrogen peroxide reacted with ethylbenzene to generate semi- and nonvolatile products [15, 33]. These products nucleated and condensed to form SOA particles after reaching their saturated vapor pressure. Thus, few particles were formed until the certain amount of ethylbenzene reacted to yield gaseous concentrations that exceed saturation vapor pressure. In the following 15~180 min, more photo-oxidation products are formed to participate in the process of gas/particle partitioning [34], and the mass concentration of SOA gradually increases. At 180 min of illumination reaction, ethylbenzene is not detected by GC-FID, while the concentration of SOA in the chamber reached the maximum value of 105 μ g/m³. Thereafter, no ethylbenzene to react to generate particles, the concentration of SOA decreased gradually due to the chamber wall effect [26-28].

For the experiments of acid NaNO₃ fine particles, the concentration of ethylbenzene SOA was obtained via subtracting the concentration of acid NaNO₃ particles from the aerosol mass concentration measured by $PM_{2.5}$ particle detector. As displayed in Fig. 1, the curve of the concentration of ethylbenzene SOA with 100 µg/m³ acidic NaNO₃ fine particles with illumination time is similar to the case of without fine particles. Nevertheless, the maximum concentration of ethylbenzene SOA in the chamber achieved



Fig. 1. The concentration of ethylbenzene SOA at different irradiation time in the absence of and presense of $100 \ \mu g/m^3$ acidic NaNO, fine particles.



Fig. 2. The maximum concentration of ethylbenzene SOA in the presence of different concentrations of acidic $NaNO_3$ fine particles.

162 μ g/m³ when illuminated for 180 min, which was 54% higher than that without fine particles. In addition, the maximum concentration of SOA produced by 1 ppm ethylbenzene and 10 ppm hydrogen peroxide in the presence of different concentrations of acidic NaNO₃ fine particles as illustrated in Fig. 2 is about 116-168 μ g/m³, which is greater than the maximum mass concentration of SOA produced without fine particles (105 μ g/m³). These demonstrated that the acidic NaNO₃ fine particles promote the formation of ethylbenzene SOA. It is worth noting that the maximum concentration of ethylbenzene SOA did not continue to increase when the concentration of acid NaNO₃ fine particles was greater than 100 μ g/m³. In order to further characterize the chemical composition of ethylbenzene SOA, UV-Vis and LC-MS measurements were performed in the following section.

Characterization of NACs in Ethylbenzene SOA with Acidic NaNO₃ Fine Particles

The OH-initiated photo-oxidation of ethylbenzene ethyl benzyl generates radical and ethyl hydroxycyclohexadienyl through hydrogen abstraction and OH addition [33, 35-36]. As displayed in Fig. 3, the ethyl benzyl radical can react with O₂ leading to the formation of phenylacetaldehyde, and the ethyl hydroxycyclohexadienyl radical reacted with O₂ to generate ethylphenol and peroxy radical. The latter undergone reactions to form glyoxal, 2-ethyl-2,4hexadienedial and other aldehydes [33, 35-36]. Also, the formed aldehydes, like phenylacetaldehyde, glyoxal can be oxidized by OH radical and O₂ and to phenylacetic acid and glyoxylic acid [6, 33]. As suggested by Fan and Zhang [37], the channel of ethyl hydroxycyclohexadienyl radical reaction with O₂ to form ethylphenol was relatively minor, indicating that carbonyls are the major constituents of ethylbenzene SOA without acidic NaNO₂ fine particles. This was confirmed by the ESI-MS and UV-Vis spectra of ethylbenzene SOA particles in the absence of fine particles displayed in Fig. 4 and Fig. 5. According to the experimental results of Carlton et al. [38], the molecular-related ions [M–H]⁻ provided the information of molecular weights of organics. The [M-H]⁻ ions of carbonyls, such as



Fig. 3. Proposed reaction mechanism for the OH-initiated photo-oxidation of ethylbenzene leading to the formation of carboxyl compounds.



Fig. 4. Negative ion mass spectra of ethylbenzene SOA particles without acidic NaNO₃ fine particles.

glyoxal (m/z 57), glyoxylic acid (m/z 73), oxalic acid (m/z 89), 4-oxo-2-hexenal (m/z 111), phenylacetaldehyde (m/z 119), 2-ethyl-2,3-epoxy-butanedial (m/z 126), phenylacetic acid (m/z 135), 2-ethyl-2,4-hexadienedial (m/z 137), 2-ethyl-2,3-epoxy-succinic acid (m/z 158) and 2-ethyl-2,4-hexadienedioic acid (m/z 169) are appeared clearly in the negative ESI-MS spectra shown in Fig. 4. Also, the peak at 205 nm of characteristic absorption of carbonyls [38] is presented in the UV-Vis spectra as displayed in Fig. 5a).

Different from the absorption peak at 205 nm of ethylbenzene SOA without fine particles, the absorption peak in the UV-Vis spectra of ethylbenzene SOA particles with 100 μ g/m³ acidic NaNO₃ fine particles shown in Fig. 5b) was significantly red-shifted to 365 nm. This peak was also observed in the UV-Vis spectra of brown carbon constituents of toluene SOA with NO_x conducted by Lin et al. [13]. They considered nitrophenols of NACs, such as nitrophenol, nitrocatechol and dinitrophenol were responsible for the



Fig. 6. Negative ion mass spectra of ethylbenzene SOA particles with $100 \ \mu g/m^3$ acidic NaNO, fine particles.

light absorption of toluene SOA with NO_x, the $n \rightarrow \pi^*$ transition in the N=O was the principal contributor to the absorption peak at 365 nm. In addition, the obtained mass spectra of ethylbenzene SOA particles with 100 µg/m³ acidic NaNO₃ fine particles shown in Fig. 6 contained the [M–H]⁻ ions of nitrophenols, such as 3-nitro-phenol (m/z 138), 3-nitro-catechol (m/z 154), 3-nitro-o-ethylphenol (m/z 166), 5-nitro-3-ethylcatechol (m/z 182), 3,5-dinitro-phenol (m/z 183) and 3,5-dinitro-o-ethylphenol (m/z 211), confirming that the 365 nm chromosphore correspond to nitrophenols.

Optical Characterization of NACs in Ethylbenzene SOA

Aerosol particles can affect atmospheric visibility through light absorption. Updyke et al. [31] and Powelson et al. [32] used the averaged mass absorption coefficient <MAC> to quantify the effect of aerosol light absorption. The total organic carbon concentration of ethylbenzene SOA with different concentrations of acid



Fig. 5. The UV-Vis spectra of ethylbenzene SOA particles a) without and b) with 100 µg/m³ acidic NaNO₃ fine particles.



Fig. 7. The averaged mass absorption coefficient (<MAC>) of ethylbenzene SOA in the presence of different concentrations of acidic NaNO, fine particles.

NaNO₃ fine particles was measured by the total organic carbon analyzer. Combined with the UV-Vis spectra of SOA collection solution in the range of 200-600 nm, the <MAC> value was calculated according to the formula (1) and (2). As shown in Fig. 7, in the absence of fine particles, the <MAC> value of ethylbenzene SOA in the range of 200-600 nm is 82 cm²•g⁻¹, which is slightly larger than that of 1,3,5-trimethylbenzene SOA particles (70 cm²•g⁻¹) measured by Updyke et al. [31]. As mentioned above, aldehydes and carboxylic acids are the principal components of ethylbenzene SOA without fine particles. These compounds only contained

C=O and C=C, did not have strong chromosphores and auxochromes [11, 39]. Therefore, ethylbenzene SOA without fine particles had weak absorbability.

The <MAC> value of ethylbenzene SOA particles increased significantly with the presence of acidic NaNO3 particles, and gradually increased with the increase of the concentration of acidic NaNO₃ particles (Fig. 7). The <MAC> value of ethylbenzene SOA in the range of 200-600 nm with 100 μ g/m³ acidic NaNO₂ fine particles is 155 cm²•g⁻¹, which is in the range of the <MAC> value of organic aerosol from biomass burning (100-700 cm²•g⁻¹) measured by Chakrabarty et al. [40], and is 89% higher than that of ethylbenzene SOA without fine particles (82 cm²•g⁻¹). This is mainly because NACs are the main components of ethylbenzene SOA with acidic NaNO₂ particles. The N=O chromosphore in NACs enhances the light absorption capacity of SOA. However, when acidic NaNO₃ fine particles exceed 100 $\mu g/m^3$, the $\langle MAC \rangle$ of ethylbenzene SOA was nearly keep constant as illustrated in Fig. 7.

Discussion

Effects of Acidic NaNO₃ Fine Particles on the NACs of Ethylbenzene SOA

The proposed model of formation mechanism of nitrophenols in the presence of acidic $NaNO_3$ fine particles is shown in Fig. 8. In the presence of acidic $NaNO_3$ fine particles, the formed gaseous products can coagulate on the aqueous surface of fine particles to form SOA before arriving its saturated vapor pressure

Fig. 8. The proposed model of formation mechanism of nitrophenols in the presence of acidic NaNO₃ fine particles.



[26]. As suggested by Vione et al. [16], the H^+ and NO_3^- can generate OH and NO_2 radicals via photolysis upon UV irradiation:

$$NO_{2}^{-}+h\nu + H^{+} \rightarrow OH+NO_{2}$$

These two radicals reacted with the coagulated phenolic products of ethylbenzene to vield NACs via the hydroxylation and nitration. As displayed in Fig. 9, OH radical can extract the hydrogen atom of phenolic hydroxyl group of o-ethylphenol to generate o-ethylphenoxy radical, which further reacted with NO, radical to form 3-nitro-o-ethylphenol and 3,5-dinitroo-ethylphenol, respectively [7, 13]. In addition, OH radical can also be added to o-ethylphenol to form 3-ethyl-catechol, which continued to react with OH and NO₂ radicals to yield 5-nitro-3-ethyl-catechol [7,13]. As suggested by Noda et al. [41] that the dealkylation of alkylbenzenes was the significant reaction path for the generation of phenolic compounds. Phenol is also produced by de-ethylation after OH radical added to the ipso position of benzene ring. Similarly, phenol reacted with OH and NO, radicals to generate 3-nitro-phenol, 3,5-dinitro-phenol and 3-nitro-catechol as outlined in Fig. 9. There was competition between the pathway of phenolic and carbonyl products from the photo-oxidation of ethylbenzene [33, 35-36]. Since the formed gaseous phenolic products immediately condensed and nitrated on the surface of acid NaNO, fine particles to form NACs, the reaction of OH radical and ethylbenzene was mainly in the direction of generating phenolic products.

Thus, NACs are the main components of ethylbenzene SOA with acid NaNO₃ fine particles.

Effects of Acidic NaNO₃ Fine Particles on the <MAC> of Ethylbenzene SOA

The acidic NaNO₃ fine particles have large specific surface area and can provide the reaction center for the formation of NACs [26]. The high concentration of acidic NaNO₂ fine particles provided more surface area for condensation and reaction of phenolic compounds. So, nitrophenols of NACs and the <MAC> of SOA increased with the concentration of acidic NaNO, fine particles. However, the concentration of ethylbenzene and hydrogen peroxide were nearly constant for every experiment, and the formed phenol and ethyl-phenol were basically unchanged. When acidic NaNO₂ fine particles were increased to a certain concentration (100 $\mu g/m^3$), all the phenol and ethyl-phenol were completely consumed, and the generation of nitrophenols would not increase with the increasing of the concentration of acidic NaNO, fine particles. Thus, the <MAC> of ethylbenzene SOA remain basically unchanged when acidic NaNO₂ fine particles exceed 100 μ g/m³ as illustrated in Fig. 7.

Compared to the experiments of Sato et al [7] and Liu et al [13], which detected NACs of nitrophenols as major light-absorption products for benzene, toluene and 1,3,5-trimethyl benzene SOA particles with NO_x , our study chemical characterized the NACs components of ethylbenzene SOA with different concentrations of



Fig. 9. Proposed reaction mechanism leading to the formation of nitrophenols.

acidic NaNO₃ fine particles, and the mass absorption coefficient averaged over 200-600 nm (<MAC>) of SOA were measured by the UV-Vis spectrophotometer and total organic carbon analyzer. From ESI-MS and UV-Vis spectra of ethylbenzene SOA, we affirmed that nitrophenols, nitrocatechols and their derivatives are the principal NACs constituents in SOA with acidic NaNO₃ fine particles. The <MAC> of ethylbenzene SOA with 100 μ g/m³ acidic NaNO, fine particles was comparable to that of biomass burning organic aerosols (BBOAs), and was greater than that of SOA without fine particles. The increase of electrophilic nitro groups strengthened the light absorption ability of SOA. As the rate generation of SOA was greater than that of BBOAs [42], the aromatic SOA particles with high concentration of acidic NaNO, fine particles would the major contributor to the climate forcing in urban atmosphere.

Conclusions

Nitro-aromatic compounds are important contributors to the light absorption by brown carbon. Aromatic compounds are the vital precursors of NACs in urban atmospheres. $\rm PM_{2.5}$ particle detector, UV-Vis spectrophotometer, LC-MS and total organic carbon analyzer were used to measure the mass concentration, NACs composition and the mass absorption coefficient averaged over 200-600 nm (<MAC>) of ethylbenzene SOA with different acidic NaNO, fine particles in this study. The obtained absorption and mass spectra showed that carbonyl compounds were the main component of ethylbenzene SOA without fine particles. Due to the large specific surface area and water absorption of acidic NaNO₃ fine particles, and the gaseous phenolic products from the photo-oxidation of ethylbenzene condensed and nitrated on the surface of the fine particles to form nitrophenols, nitrocatechols and other NACs, thus significantly promoting the formation of ethylbenzene SOA. The <MAC> value of SOA gradually increased with the increasing of the concentration of acidic NaNO₃ particles, and the \leq MAC \geq of SOA with 100 µg/m³ acidic NaNO₃ particles is close to that of biomass burning organic aerosols. However, the <MAC> value of SOA was measured by off-line method, optical parameters such as extinction and scattering coefficient of SOA should be detected on-line in the subsequent experiments.

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Conflict of Interest

The authors declare no conflict of interest.

Reference

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